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# Current and future nitrous oxide emissions from African agriculture

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Most emission estimates of the greenhouse gas nitrous oxide ( $N_2O$ ) from African agriculture at a continental scale are based on emission factors, such as those developed by the IPCC Guidelines. Here we present estimates from Africa from the EDGAR database, which is derived from the IPCC emission factors. Resulting estimates indicate that  $N_2O$  emissions from agriculture represented 42% of total emissions from Africa (though that rises to 71% if all savannah and grassland burning is included), or roughly 6% of global anthropogenic  $N_2O$  emissions (or 11% including burning). Emissions from African agriculture are dominated by grazing livestock; 74% of agricultural  $N_2O$  excluding biomass burning was from paddocks, ranges, and pasture. Direct soil emissions represent 15% of agricultural emissions; substantial changes in direct emissions from North Africa helped drive a 47% continental increase in direct soil emissions from 1970 to 2005. Future trends based on the Millennium Ecosystem Assessment scenarios indicate that agricultural  $N_2O$  emissions may double in Africa by 2050 from 2000 levels. Any regional or continental estimates for Africa are, however, necessarily limited by a paucity of direct measurements of emissions in sub-Saharan agro-ecosystems, and the heavy reliance on emission factors and other default assumptions about nitrogen cycling in African agriculture. In particular, a better understanding of livestock-related N inputs and  $N_2O$  emissions will help improve regional and continental estimates. As fertilizer use increases in sub-Saharan Africa, emission estimates should consider several unusual elements of African agriculture: farmer practices that differ fundamentally from that of large scale farms, the long history of N depletion from agricultural soils, seasonal emission pulses, and emission factors that vary with the amount of N added.

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## Introduction

Agriculture is a major source of atmospheric nitrous oxide ( $N_2O$ ), one of the three most important greenhouse gases [1]. In general, emissions of  $N_2O$ , which are produced not only during denitrification but also during nitrification [2], are most strongly influenced by nitrogen (N) availability and soil moisture, and to a lesser extent by factors such as temperature, pH, and the availability of reduced carbon [3]. In agricultural production systems, inputs of inorganic and organic N provide a source of ammonium and nitrate for nitrification and denitrification, while higher soil moisture levels provide the anoxic conditions required for denitrification, and slow the diffusion of nitric oxide (NO) through soils, creating further opportunities for its reduction to  $N_2O$ . Application of N fertilizers and animal production accounts for about 60% of the global  $N_2O$  emissions from agriculture [4]. Since 1970,  $N_2O$  emissions have grown mainly as a result of increased use of mineral fertilizers worldwide [5], but in Africa, N input via synthetic N fertilizers is among the lowest in the world, representing 3% of the world fertilizer consumption [6]. The African continent, and sub-Saharan Africa (SSA) in particular, makes a small contribution to global greenhouse gas emissions from agriculture: Africa has been estimated to produce only 15% of global agriculture soil  $N_2O$  emissions and 3% of emissions from animal production systems [7].

In part as a consequence of extremely low fertilizer inputs in Africa (an average of 8 kg nutrients  $ha^{-1}$  compared to 100 kg N  $ha^{-1}$  in the United States, and 220 kg N  $ha^{-1}$  in China) [8••], N removal through crop harvest and erosion over the last several decades has led to negative nutrient balances, declining grain yields, and consequently to insufficient food intake to meet dietary protein and energy requirements [9]. Currently 250 million people in SSA, 30% of the total African population, are chronically malnourished [10]. A greater input of N to the soil is urgently needed to satisfy the basic needs of an increasing African population. This increase in N inputs is likely to be the largest perturbation to the N cycle in Africa in coming decades. One inevitable impact of increasing N inputs is the growth of direct and indirect  $N_2O$  emissions, although the magnitude of these emissions may be influenced by environmental conditions (such as the retention of nitrate on anion exchanges sites in subsoils of certain soil types in SSA) or management practice (such as spatial and temporal targeting of fertilizer application to plant demand to increase rates of N uptake and reduce losses to the environment) [11–13].

Several recent successes have paved the way for a dramatic expansion of fertilizer use in SSA, which has

gained momentum with Kofi Annan's call for an African Green Revolution. In Malawi, a national subsidy program providing smallholder farmers with vouchers for fertilizer and improved seed has consistently resulted in a doubling of national grain yields, and has helped the country shift from a maize production deficit to a surplus [14]. The Millennium Villages Project, which also promotes subsidized fertilizer use, has also demonstrated that in combination with improved maize varieties and extension services, increasing N application (post-intervention application rates for maize ranged from 45 to 130 kg N ha<sup>-1</sup> in the different villages) can quickly double and triple maize yields across SSA [15,16]. The implications of this increased application of N inputs to agricultural systems in Africa for N<sub>2</sub>O emissions are still only partially understood. For example, there are no published studies from Africa examining the response function relating N<sub>2</sub>O emissions to incremental increases in N inputs (e.g. [17]).

Current estimates of N<sub>2</sub>O emissions from African agriculture at the national, regional, and continental scale are mostly based on the IPCC Guidelines [7,18], which implicitly ignore important characteristics of African soils and management practices of smallholder farmers, both of which could alter emissions substantially. Some efforts have been made to improve the estimates by applying statistical and process-based models [19,20,21]. However, these estimates are often hindered by a lack of the data needed as basic inputs for these models, and by the relatively small number of emission measurements made in SSA. In particular, the available weather station observations in sub-Saharan Africa (e.g. [22]) are plagued with large and numerous recording gaps, and current soil maps are generally based on sparse and often incomplete soil profile measurements, particularly in comparison to western countries [23]. Because N availability and soil moisture are the major determinants of N oxide emissions [24–26] and are related to basic soil properties, these data

shortfalls make regional estimates of N<sub>2</sub>O emissions challenging.

In this paper we review some existing estimates of the current and future trends in N input and N<sub>2</sub>O emissions from African agriculture. Next we discuss limitations of the estimation methods for Africa and what is needed to address these limitations.

## Trends in N inputs and nitrous oxide emissions

### Current sources and rates

Agriculture is a source of N<sub>2</sub>O through direct and indirect emissions from soils. First of all, N<sub>2</sub>O production by bacteria in agricultural fields generally increases from soils where N is added, referred to as *direct* emissions from agriculture. In addition, N losses from agricultural systems through leaching, runoff, erosion, or emissions of ammonia or NO may result in N<sub>2</sub>O production at remote sites, which is generally referred to as *indirect* emissions.

We present N<sub>2</sub>O emissions from Africa as the sum of national N<sub>2</sub>O inventories for 1970 and 2005 as compiled in the Emission Database for Global Atmospheric Research (EDGAR) [27,28]. The EDGAR database follows the 2006 IPCC Guidelines and uses a consistent set of emission factors for N<sub>2</sub>O emission estimates from agriculture, energy, transport and industries. On the basis of these calculations, Africa contributed 15.5% to the total global anthropogenic emissions of N<sub>2</sub>O in the year 2005 (1500 Gg N<sub>2</sub>O). Agriculture contributed an estimated 42% of the African N<sub>2</sub>O emissions, though that total rises to 71% if all emissions from grassland and savannah burning is included (Table 1).

### Livestock sources

The largest source of agricultural emissions was livestock; combined emissions from paddocks, ranges, and pastures accounted for 74% of all African agricultural

**Table 1**

**Regional N<sub>2</sub>O emissions from agriculture soil, indirect agriculture emissions, manure management, pasture/range/paddock, and savannah and grassland burning in Northern Africa and sub-Saharan Africa (Eastern, Western and Southern) for 2005 and global N<sub>2</sub>O emissions for the year 1970 and 2005 (EDGAR v.4.1). Regional decreases in Western and Southern Africa are largely due to decreases in savannah burning. Increases in Eastern Africa are primarily due to an increase in emissions from pasture/range/paddock of over 50%.**

Region*	Total agric. N <sub>2</sub> O (Gg yr <sup>-1</sup> )		N <sub>2</sub> O emissions by category in 2005 (Gg yr <sup>-1</sup> )				
	1970	2005	Direct soil	Indirect	Manure management	Pasture/range/paddock	Savannah and grassland burning
Northern	36	81	32	11	1	38	0
<i>Sub-Saharan Africa</i>							
Eastern	211	343	16	16	3	192	115
Western	431	348	19	14	6	120	190
Southern	366	287	28	9	3	97	150
Global estimates	4230	6470	2610	858	324	1900	736

Source: [28].

N<sub>2</sub>O emissions, excluding savannah and grassland burning; emissions from on-farm manure management (including handling, storage, and application) accounted for an additional 3% (Table 1).

Emissions from (less intensively managed) livestock holdings in the tropics are generally expected to be lower than those from (intensively managed) holdings in temperate ecosystems [29], and there may be reasons to be wary of the EDGAR estimates. Emissions of N<sub>2</sub>O from animal production are mainly determined by manure management system, N excretion rate, animal density, animal type, and the percentage N lost as N<sub>2</sub>O (which is independent of region) [29]. The large emissions from pasture compared to manure management is largely due to EDGAR's use of the 2006 IPCC guidelines for allocation of manure to different management systems: 83–100% of manure is assigned to the paddock/range/pasture category in Africa, but only 1–5% is assigned to manure management.

There may be a need to re-evaluate the N excretion rates used by EDGAR. Though the N content of manure in the tropics is commonly reported to be less than 1/3 as much as in temperate agroecosystems [30], the default IPCC factors for estimating N excretion rates in Africa are the highest or second highest in the world for most categories of livestock [18]. For example, the most recent IPCC guidelines estimate that 0.60 kg N day<sup>-1</sup> is lost per 1000 kg of dairy cattle in Africa but only 0.35–0.48 kg N day<sup>-1</sup> in Western Europe. Recent studies suggest that N excretion rate estimates are in need of revision: cattle in Sweden excreted N at rates 2–6 times higher than cattle in Kenya or Mali [30]. The 2006 IPCC default rates are also a stark reversal of the 1996 guidelines, in which dairy cattle in Western Europe and North America excreted an estimated 100 kg N head<sup>-1</sup> yr<sup>-1</sup> compared to 60 kg N head<sup>-1</sup> yr<sup>-1</sup> in Africa, though these may still be too high for Africa [30,31].

Livestock density in TLU km<sup>-2</sup> tends to be low in Africa, ranging from 0 to 5 in pastoral system in arid/semi-arid regions to 35–55 TLU km<sup>-2</sup> in temperate/tropical highlands [30]; densities are much higher in other parts of the world (e.g. 2 sheep ha<sup>-1</sup> [32]). The low densities in Africa would be expected to contribute to relatively low emission rates per unit area (though livestock-related point sources (e.g. corrals) can become the dominant source of N<sub>2</sub>O [32], but EDGAR does not address this level of spatial detail). But when calculated on a per unit area basis using FAO land-use classifications, the EDGAR mean emission rates from African pasture (0.16 kg N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup>) are roughly equal to emissions for North American and European pasture (0.16 and 0.17 kg N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup>, respectively) [33]. In contrast, observed emissions from intensively managed unfertilized dairy pasture in western countries found emission rates that

ranged from 1 to 2 orders of magnitude higher than these averages [34,35], though there are others that range from roughly equal to these averages to 4 times higher [36]. Comparable measurements have not been published for sites in SSA.

On-farm manure management may make a small contribution to African agricultural N<sub>2</sub>O emissions for several reasons. Application rates can be quite low because cattle may graze widely, making manure collection challenging compared to more intensive livestock operations, and farmers in some regions rarely use manure as fertilizer [12,30,37]; Liu *et al.* estimate that manure represents 15% of N inputs to crops in Africa [38]. Once collected, substantial N can be lost from manure during storage, leading to lower rates of N application to fields [30,39]. The losses during storage are primarily ammonia; N<sub>2</sub>O emissions can be 5–10 times smaller than the ammonia emissions (which can lead to indirect N<sub>2</sub>O emission), as observed for dung storage in Niger [40].

#### Other sources and sinks

Savannah burning and grassland fires represent an additional substantial source of N<sub>2</sub>O emissions in Africa, roughly equivalent to emissions from livestock, much of which can be attributed to land preparation (Table 1). Burning of agricultural waste is the smallest source of N<sub>2</sub>O from agriculture in Africa, which at 1.54 Gg is an order of magnitude smaller than emissions from manure management.

According to the EDGAR database, direct soil emissions from African agriculture compromise 15% of the total agriculture N<sub>2</sub>O emissions excluding burning, and have been increasing linearly, from 42 Gg in 1970 to 95 Gg in 2005, largely from increases in Northern Africa (Table 1). The relatively small direct emissions can be attributed to the extremely low rates of inorganic fertilizer use in most of sub-Saharan Africa [8\*\*]. Liu *et al.* estimate that of the anthropogenic N inputs in African cropland, 60% is from biological N fixation (BNF; 2.9 Tg yr<sup>-1</sup>) and mineral N fertilizers (2.2 Tg yr<sup>-1</sup>), 10% from atmospheric deposition, 15% from animal manure and 12% from crop residue [38]. The IPCC Guidelines indicate that N<sub>2</sub>O emissions from soybean production are negligible, but there are indications that cultivation of other legumes used for enhancing soil N (e.g. cover crops, tree fallows) may result in higher N<sub>2</sub>O emissions than emissions from applications of mineral fertilizers [24,41]. More complete studies are needed with mineral fertilizers and organic inputs on different soil types and different rates to determine the relative roles of mineral and organic fertilizers as well as BNF in N<sub>2</sub>O emissions, especially as N inputs begin to increase (e.g. [24,25,42]).

In Africa, most N added to the landscape from any source is estimated to be lost through nitrate leaching [43],

making rivers and riparian zones a potentially important indirect source of N<sub>2</sub>O emissions [44<sup>\*</sup>]. Most studies indicate that indirect agricultural N<sub>2</sub>O emissions may account for about 25% of the global agricultural emissions [40–43,44<sup>\*</sup>,45–47]. The estimate by EDGAR, however, suggests that African indirect N<sub>2</sub>O emissions accounted for just 8% of the total African agricultural emissions in 2005.

Some soils can also act as sink of N<sub>2</sub>O emissions at times, especially in wet and water logged conditions, though not enough to offset emissions at a regional scale [48,49]. National inventories for African countries and the EDGAR database do not consider soils as N<sub>2</sub>O sink because our understanding of the sink strength of African soils is based on very few studies. One study in Mozambique [48] assumed soil uptake rate at 0.0087 kg N ha<sup>-1</sup> yr<sup>-1</sup> and indicates that 7–18% of total N<sub>2</sub>O emissions could be consumed through soil processes.

A description and quantification of uncertainty in EDGAR 4.0 is currently in preparation, but has not yet been published [28]. Uncertainties in the emission factors for grazing animals and for indirect emissions have been identified to have a large effect on global N<sub>2</sub>O emission estimates; for example, the estimated range of N lost from manure as N<sub>2</sub>O spans nearly 2 orders of magnitude [49], and IPCC's uncertainties around estimates of N excretion rates are 50% [18]. The dominant role played by livestock in Africa N<sub>2</sub>O emissions highlights the importance of reducing these uncertainties for improving continental emission estimates.

#### Future trends in increasing N inputs to African agriculture

Increasing N inputs to intensify agricultural production and satisfy demands for food in sub-Saharan Africa should result in increased N<sub>2</sub>O emissions [7,10]. Here we present some projections of N inputs and the implications for N<sub>2</sub>O emissions to 2050 from the Millennium Ecosystem Assessment (MEA) [50,51]. The four MEA scenarios include: Global Orchestration (GO), Order from Strength

(OS), Techno-garden (TG) and Adapting Mosaic (AM). They differ with respect to whether globalization is assumed to be the dominant trend (as in GO and TG), or regionalization (as in OS and TG). In globalized worlds, the worldwide economies are more open than in regionalized worlds. Moreover, the scenarios differ in the assumed approach towards ecosystem management, which may be pro-active (as in AM and TG), or reactive (as in GO and OS). The scenarios with a pro-active approach towards ecosystem management either focus on large-scale technical solutions (in the globalized TG world), or more on small-scale, local solutions (in the regionalized AM world). This has consequences for the position of small holders and large-scale producers in agriculture.

In all four scenarios the African population and economies grow [52,53]. In 2000, 796 million people lived in Africa with an average GDP of 745 1995US\$ person<sup>-1</sup> yr<sup>-1</sup>. The growth rates differ among scenarios. In GO, for instance, it is assumed that the population in 2050 is 80% larger than in 2000 (1.4 billion) while per capita has GDP tripled (2993 1995US\$ person<sup>-1</sup> yr<sup>-1</sup>). In OS, on the other hand, the population increases more quickly (to 2 billion) while per capita GDP doubles (1488 1995US\$ person<sup>-1</sup> yr<sup>-1</sup>).

Among other assumptions, the future scenarios assume changes in the human diets towards more meat consumption; in Africa, consumption of meat (animal protein) per capita increases by a factor of 2–3. The scenarios also assume an increase in the share of pork and poultry in total livestock production [52]. As a result of these assumptions, animal production as well as manure excretion increases in Africa. In addition, agricultural changes are envisaged that result in a more efficient use of fertilizers.

The MEA scenarios were quantitatively interpreted for the different continents up to the year 2050 (Table 2) [52]. Balances of N, estimated as the total inputs (fertilizer, manure, human N, N<sub>2</sub> fixation, and N deposition) minus N removed through harvest and grazing, increase

**Table 2**

**N balances for the world and for Africa. Results are shown for 1970 and 2000, and for Millennium Ecosystem Assessment Scenarios for the year 2050: Global Orchestration (GO), Order from Strength (OS), Technogarden (TG) and Adapting Mosaic (AM).**

Year	Global N balance (Gg N yr <sup>-1</sup> )	African N balance (Gg N yr <sup>-1</sup> )	African N balance (kg N ha <sup>-1</sup> yr <sup>-1</sup> ) <sup>*</sup>
1970	101 000	12 700	17
2000	157 000	18 200	18
2050 GO	231 000	42 400	26
2050 OS	211 000	36 800	23
2050 AM	154 000	25 600	20
2050 TG	172 000	37 000	22

Source: [52].

<sup>\*</sup> For readers accustomed with other units: 1000 kg N km<sup>-2</sup> = 10 kg N ha<sup>-1</sup> yr<sup>-1</sup>.

globally in all scenarios [52]. For most regions, however, the N balances decrease in the scenarios with a proactive attitude (TG and AM). Africa is an exception to this where there is an increase in all scenarios. These trends in African N balances result from the assumption that land degradation by nutrient depletion is reduced or halted through increased additions of N fertilizers and manures to agricultural land [52]. There are also increases in N balances per hectare of land in Africa, and they are relatively high compared to other world regions. In some parts of North Africa with intensive agriculture, N balances are comparable to those in industrialized countries.

The net effect of these assumed trends is that in Africa, the N inputs to agricultural fields (from inorganic fertilizer and manure) are 1.5–6 times the 2000 level in 2050. The N balances in 2050 are at least twice the 2000 level. This is an important change. Currently, N depletion is a problem in large parts of Africa. According to the MEA scenarios, active nutrient depletion will be less important in the future as a result of increased N inputs. The increasing N inputs also imply that the emissions of N<sub>2</sub>O will increase during the 21st century. Assuming a linear relation between N inputs and N<sub>2</sub>O emissions (as in most emission inventories), we may argue that N<sub>2</sub>O emissions from African agriculture roughly double during the coming four decades, depending on the scenario. Indeed, the Integrated Model to Assess the Global Environment (IMAGE) model projects an increase in land use related N<sub>2</sub>O emissions from Africa, increasing from 300 Gg in 1970, to 800 Gg in 2010 and 1200 in 2050 [54].

### Considerations in applying different emission estimation approaches in Africa

#### Limitations of the IPCC Emission Guidelines

Though of great utility, there are important limitations of the IPCC Guidelines for application to N<sub>2</sub>O emissions from African agriculture, as we touched on above with respect to livestock. The IPCC Guidelines provide a model for estimating N<sub>2</sub>O fluxes based on calculations of mean emission rates: as a default, 1–2% of added N is assumed to be emitted as direct N<sub>2</sub>O emissions, and roughly 1% is assumed to be indirect as N<sub>2</sub>O emissions from N that has been transported off-farm. Top-down estimates based on changes in atmospheric N<sub>2</sub>O suggest that the IPCC Guidelines underestimate actual emissions, and indicate instead that roughly 4% of atmospheric N newly fixed through natural or industrial processes is lost as N<sub>2</sub>O [55]. However, evaluation against the atmospheric record suggests that using a constant emission rate does not adequately reflect the complexities of biogenic N<sub>2</sub>O fluxes, even for the purpose of global estimates [56].

Recent changes in our understanding of the relationship between N inputs and N<sub>2</sub>O emissions at the plot scale further highlight the limitations of applying emission

factors widely, particularly to areas where much lower rates of N are applied. Mounting evidence suggests that the rate of soil N<sub>2</sub>O emission varies nonlinearly in response to increasing N additions, with a smaller percentage of added N lost as N<sub>2</sub>O at lower rates of fertilization and a larger percentage lost at higher rates: experimental and on-farm studies in temperate agro-ecosystems have demonstrated clear thresholds in N<sub>2</sub>O emission responses to N inputs, with N<sub>2</sub>O emissions increasing more rapidly once N inputs exceed plant demand, typically above 100–135 or 187 kg N ha<sup>-1</sup> in high-input mechanized agro-ecosystems [17,57,58]. In crop trials, inorganic fertilizer is typically applied to maximize yields, frequently at rates of 100–150 kg ha<sup>-1</sup>, much more than the current 8 kg ha<sup>-1</sup> average, or even the 70–80 kg ha<sup>-1</sup> rates recommended by many countries in sub-Saharan Africa. If the nonlinear responses observed in temperate agro-ecosystems apply in Africa, one would expect a smaller percentage of added N to be lost as N<sub>2</sub>O than is observed in the studies used to develop the IPCC emission factors. Experimental work in Africa has largely not included the multiple levels of N inputs needed to identify a non-linear N<sub>2</sub>O response, except in studies where changes in the amount of N applied is confounded with the form of N applied (e.g. [24,59–61]), studies where inputs were applied only at high rates from 120 to 3800 kg N ha<sup>-1</sup> (thus excluding the entire lower portion of the typical response curve, e.g. [60,61]), and studies where rates of N input were not reported at all (e.g. [42,62]). The one example with measurements of N<sub>2</sub>O emissions from multiple levels of a single type of (inorganic) fertilizer included only three levels of fertilizer, limiting insights into the linearity of the fertilizer/emission relationship [25].

Emission rates from cropped soils can also be influenced by how well fertilizer applications are placed and timed to plant demand. IPCC emission factors may be substantially influenced by fertilizer application methods in studies used for the emission factor calculations. Small-holder farmers in sub-Saharan Africa apply fertilizer in a split application (to meet the timing of plant demand), and place the fertilizer strategically: in the seeding hole at planting, and immediately adjacent to each plant at topdressing. These practices are intended to synchronize N availability with plant N demand to increase N use efficiency, and may further reduce the percentage of N applied that is lost as N<sub>2</sub>O and in other forms.

An additional consideration in applying IPCC emission factors to Africa is the fact that they were developed based on existing measurements of NO and N<sub>2</sub>O emissions, few of which were conducted in tropical ecosystems, and hardly any in Africa. Sub-Saharan Africa is particularly poorly represented, with all the studies used by the IPCC conducted in South Africa and Zimbabwe [45,63], and only one — a study of NO emissions — conducted in an agricultural site [26]. Only one study of N<sub>2</sub>O emissions

from sub-Saharan Africa was used, from an unfertilized natural savannah [64].

In general, few measurements have been made of N<sub>2</sub>O emissions from soils in sub-Saharan agricultural systems, and many of those that have been made provide an incomplete picture of the effects of various types of agricultural management on trace N gas emissions, or the flux response to increasing N inputs. Among studies directly measuring N<sub>2</sub>O emissions in sub-Saharan agroecosystems, the majority have examined the effects of leguminous tree fallows on N<sub>2</sub>O emissions (e.g., [60]), including several lab studies (e.g. [65]), while others have examined different fertilizer options [25], compared rain-fed savannah to irrigated *Acacia* plantation [66], or examined N<sub>2</sub>O fluxes from high-input urban vegetable gardens [58].

Emission factors can be an effective way to compare N<sub>2</sub>O losses from agriculture in different regions, but estimating annual emission factors for these studies is challenging, as only three studies in Africa have conducted measurements over the course of an entire year [25,61,62]. Measurements that do not extend beyond the growing season present particular problems in seasonally dry natural systems, where emissions are generally believed to be low during the dry season, and where pulses associated with early wet-season precipitation events can be responsible for a substantial proportion of annual N trace gas fluxes [67]. Emissions of N<sub>2</sub>O have approached 60  $\mu\text{g N m}^{-2} \text{h}^{-1}$  from a dry-season range of  $-0.2$  to 3.4  $\mu\text{g N m}^{-2} \text{h}^{-1}$  in Malian cereal fields [62], and exceeded 150  $\mu\text{g N m}^{-2} \text{h}^{-1}$  during early season pulses in a Burkina Faso savannah; if sustained over a 24-hour period, this one-day pulse would represent over 5% of annual N<sub>2</sub>O emissions from the savannah [25]. Consequently, studies in which measurements are not frequent enough to capture pulses, in which the frequency and timing of measurements result in mean flux rates that are disproportionately influenced by measurements of pulse emissions, or in which the sampling period excludes the dry season or the beginning of the rainy season are unlikely to accurately describe annual emissions. But assuming that periods measured are representative of emission rates throughout the rest of the year, emission factors are generally at the low-end of the 0.0–7.8% range observed in studies around the world [68], with less than 1% of added N estimated to be emitted as N<sub>2</sub>O in many cases (e.g. [24,25,42,59,60]). There are exceptions, however, which include two of the three studies that conducted year-round measurements. In high-input vegetable gardens (473–3816 kg N ha<sup>-1</sup>), losses ranged from 1.6% to 5.2% of added N annually [61], and in a Pearl millet field in Mali, 4.1% of the N added in a 50 kg urea treatment was lost as N<sub>2</sub>O within the first year [62]. In a study on agroforestry in Kenyan maize fields, additions of 115 kg N ha<sup>-1</sup> and 214.6 kg N ha<sup>-1</sup> from leguminous

fallows resulted in more than 1.49% and 1.82% of the added N being lost as N<sub>2</sub>O over 84 days, though higher rates of organic N additions resulted in smaller emission factors of 0.45–0.56% [60].

Other soil factors in Africa may affect emissions. The nitrate retention capacity characteristic of the variable charge clays found in many soils in Africa may reduce N<sub>2</sub>O emissions compared to soils without this property: if nitrate is retained in the subsoil rather than leached to groundwater (e.g. [13]), it could be made effectively unavailable for denitrifiers. The long history of nutrient depletion and loss of organic matter of agricultural soils in sub-Saharan may also make for different emission responses to N inputs than in temperate or well-managed systems [62]. There are few studies expressly investigating the effects of depleted soils on gas emissions, but there are several lines of speculation: it is possible that plant and microbial uptake may be more complete in soils where existing nutrient pools are limited; alternatively, it is conceivable that these soils lack the capacity to effectively retain N (e.g. lacking microbial populations that are capable of immobilizing large quantities of N), and may result in greater losses. It is also possible that denitrification in low SOC soils may proceed more completely to N<sub>2</sub> with additions of labile C and N [62], or that soils with low SOC (and lower capacity for water retention) may experience greater nitrate leaching rates. All these factors make predictions or estimations of emissions with currently available data difficult for soils with these characteristics.

#### Applying statistical and process-based models in the African context

Many of the limitations of the IPCC default values for N<sub>2</sub>O emissions also apply to other emission estimation approaches such as statistical and process-based models of global trace N gas emissions. Process-based models have been applied at global scales to estimate N<sub>2</sub>O emissions from agriculture, including emissions from Africa. For example, DAYCENT has been used to evaluate N<sub>2</sub>O emissions from maize, wheat, and soybean under management practices for reducing greenhouse gas emissions [69], while the Terrestrial Ecosystem Model was used to evaluate the impacts of different scenarios of biofuel use on emissions [19<sup>••</sup>]. Simple emission factors have also been applied, and speak to the potential for agricultural intensification to prevent the extension of agriculture into natural habitat, resulting in lower net greenhouse gas emissions [70]. Because the value of these models is dependent on the data driving them, the poor availability of many data for Africa presents a serious challenge; for example, the Terrestrial Ecosystem Model simulations of Africa are based on parameterizations from South America (David Kicklighter, personal communication). Ongoing efforts at constructing a high-resolution digital soil map of Africa are an important step forward [71], and would be made even more valuable to modelers if paired

with increased access to weather station observations and improved, spatially explicit information on agricultural management practices.

## Conclusions

In this paper we reviewed estimates of current and future N<sub>2</sub>O emissions from agriculture in Africa. These emissions have been increasing over the past thirty years, and future scenarios project this trend to continue with increasing N inputs to agricultural land. These estimates of N<sub>2</sub>O are based on the IPCC guidelines, and though they provide an initial estimate, the current default values and assumptions of the IPCC are unlikely to accurately estimate emissions in Africa. Overall, it is evident that there are important knowledge and data gaps that must be addressed to improve current and future estimates of N<sub>2</sub>O emissions, with improvements in estimating the various livestock contributions to emissions being of particular importance for a region where fertilizer use has been historically low. However, improved estimates of direct soil emission from fertilizer use will be increasingly important for estimating continental emissions if sub-Saharan Africa undergoes the change in land-use and land management associated with a burgeoning Green Revolution.

Although formal uncertainties for the EDGAR database are still being prepared, we have identified some basic uncertainties in the African N<sub>2</sub>O budget related to a variety of factors: the size and nature of N inputs (which is also dependent on estimates of cropland area, crop management, and livestock populations and densities, each of which is plagued with its own uncertainties), the basic relationship between N inputs and N<sub>2</sub>O emissions (particularly with the low rates of N currently applied in the continent), the response of N<sub>2</sub>O emissions on N-depleted soils and soils with high potential to retain nitrate on variable charged clays, and perhaps the potential importance of soil sinks. These uncertainties can only be reduced by increasing the number of field measurements in Africa. Efforts such as the African Soil Information Service's Digital Soil Map of Africa project will provide some essential information for modelers, but greater effort at improving the availability of existing data (e.g. weather observations) and filling critical data and knowledge gaps is essential. Given the relatively small number of existing studies for the continent, coordinated efforts among African and international scientists are needed to improve estimates of N<sub>2</sub>O emissions from the continent, as well as to improve a more general understanding of N dynamics in Africa. An interdisciplinary hierarchical approach that consists of strategically designed measurements, observations, experiments, and simulations at plot, farm, landscape, watershed, regional, and continental scales can be a powerful approach to understanding the flows and cycling of N from agriculture and livestock in African environments.

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## References and recommended reading

Papers of particular interest, published within the period of review, have been highlighted as:

- of special interest
  - of outstanding interest
1. Forster PV, Ramaswamy P, Artaxo T, Berntsen R, Betts DW, Fahey J, Haywood J, Lean DC, Lowe G, Myhre J *et al.*: **Changes in atmospheric constituents and in radiative forcing.** In *Climate Change 2007: The Physical Science Basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change.* Edited by Solomon S, Qin D, Manning M, Chen Z, Marquis M, Averyt KB, Tignor M, Miller HL. Cambridge University Press; 2007:129-234.
  2. Firestone MK, Davidson EA: **Microbial basis of NO and N<sub>2</sub>O production and consumption in soils.** In *Exchange of Trace Gases Between Terrestrial Ecosystems and the Atmosphere.* Edited by Andreae MO, Schimel DS. John Wiley & Sons; 1989:7-22.
  3. Davidson EA, Verchot LV: **Testing the hole-in-the-pipe model of nitric and nitrous oxide emissions from soils using the TRAGNET database.** *Global Biogeochemical Cycles* 2000, **14**:1035-1043.
  4. Syakila A, Kroeze C: **The global N<sub>2</sub>O budget revisited.** *Greenhouse Gas Measurement and Mitigation* 2011, **1**:17-26.
  5. IPCC: *Fourth Assessment Report on Climate Change Impacts, Adaptation and Vulnerability for Researchers, Students, Policymakers.* Cambridge University Press; 2007.
  6. FAO: **Food and Agriculture Organization of the United Nations.** FAOSTAT-Agriculture Database; <http://faostat.fao.org/> (accessed in April 2011).
  7. *Global Anthropogenic Non-CO<sub>2</sub> Greenhouse Gas Emissions: 1990–2020.* Washington, DC: US-EPA: United States Environmental Protection Agency; 2006.
  8. Vitousek PM, Naylor R, Crews T, David MB, Drinkwater LE, Holland E, Johnes PJ, Katzenberger J, Martinelli LA, Matson PA *et al.*: **Nutrient imbalances in agriculture development.** *Science* 2009, **324**:1519-1520 This paper shows that nitrogen and phosphorus addition to agriculture systems in different countries range from inadequate (Kenya) to excessive (United States and China). Both of these cause serious environmental and health challenge..
  9. Sanchez PA: **Soil fertility and hunger in Africa.** *Science* 2002, **295**:2019-2020.
  10. FAO: *The State of Food Insecurity in the World.* Food and Agriculture Organization of the United Nations; 2010.
  11. Van Beek CL, Meerburg BG, Schiils RLM, Verhagen J, Kuikman PJ: **Feeding the world's increasing population while limiting climate change impacts: linking N<sub>2</sub>O and CH<sub>4</sub> emissions from agriculture to population growth.** *Environmental Science and Policy* 2010, **13**:89-961.
  12. Batterbury S, Warren A: **The African Sahel 25 years after the great drought: assessing progress and moving towards new agendas and approaches.** *Global Environmental Change* 2001, **11**:1-8.
  13. Mekonnen KR, Buresh J, Coe R, Kipling KM: **Root length and nitrate under Sesbania sesban: vertical and horizontal distribution and variability.** *Agroforestry Systems* 1998, **42**:265-282.
  14. Denning G, Kabambe P, Sanchez P, Malik A, Flor R, Harawa P, Nkhoma P, Zamba C, Banda C, Magombo C *et al.*: **Input subsidies to improve smallholder maize productivity in Malawi: towards and African green revolution.** *Public Library of Science Biology* 2009, **10**:2-7.

15. Sanchez PA, Palm C, Sachs J, Denning G, Flor R, Harawa R, Jama B, Kiflemariam T, Konecky B, Kozar R *et al.*: **The African millennium villages.** In *Proceedings National Academy of Sciences of United States of America* 2007, **104**:16775-16780 The paper describes results of the Millennium Villages Project in rural Africa, where science-based interventions have helped to reduce malaria, met caloric requirements, generate crop surplus..
16. Nziguheba G, Palm CA, Berhe T, Denning G, Dicko A, Diouf O, Diru W, Flor R, Frimpong F, Harawa R *et al.*: **The African Green Revolution: results from the Millennium Villages Project.** *Advances in Agronomy* 2010, **109**:75-115.
17. McSwiney CP, Robertson GP: **Nonlinear response of N<sub>2</sub>O flux to incremental fertilizer addition in a continuous maize (Zea mays L.) cropping system.** *Global Change Biology* 2005, **11**:1712-1719 The paper found a clear non-linear effect of incremental increases in N inputs on emissions of nitrous oxide, demonstrating that emissions increase more rapidly above certain key thresholds of N inputs. Among other things, these results suggest that the use of a single emission factor for all levels of N inputs may overestimate emissions at the input levels typical of sub-Saharan Africa..
18. IPCC: In *Guidelines for National Greenhouse Gas Inventories of the International Panel on Climate Change.* Edited by Eggleston S, Buendia L, Miwa K, Ngara T, Tanabe K. Hayama, Japan: Institute for Global Environmental Strategies; 2006. on behalf of IPCC, National Greenhouse Gas Inventory Program.
19. Melillo JM, Reilly JM, Kicklighter DW, Gurgel AG, Cronin TW, Paltsev S, Felzer BS, Wang X, Sokolov AP, Schlosser A: **Indirect emissions from biofuels. How important?** *Science* 2009, **326**:1397-1399 The paper examined direct and indirect effects of possible land-use changes from an expanded global cellulosic bioenergy program on greenhouse gas emissions including African continent. The model predicts an increase in fertilizer use and nitrous oxide emissions, which will be more important than carbon losses in terms of warming potential..
20. Stehfest E, Heistermann M, Pries J, Ojima DS, Alcamo J: **Simulation of global crop production with the ecosystem model DayCent.** *Ecological Modelling* 2007, **209**:203-219.
21. Grote R, Lehmann E, Brummer C, Bruggemann N, Szarzynski J, Kunstmann H: **Modeling and observation of biosphere-atmosphere interactions in natural savannah in Burkina Faso, West Africa.** *Physics and Chemistry on the Earth* 2009, **34**: 251-160.
22. NCDC: *COOP Data/Record of Climatological Observations.* Federal Building, 151 Patton Avenue, Asheville, NC: National Climatic Data Center; 2008.
23. Gijsman AJ, Thornton PK, Hoogenboom G: **Using the WISE database to parameterize soil inputs for crop simulation models.** *Computers and Electronics in Agriculture* 2007, **56**:85-100.
24. Baggs EM, Chebii J, Ndufa JK: **A short-term investigation of trace gas emissions following tillage and no-tillage of agroforestry residues in western Kenya.** *Soil & Tillage Research* 2006, **90**:69-76.
25. Brummer C, Bruggemann N, Butterbach-Bahl K, Falk U, Szarzynski J, Vielhauer K, Wassmann R, Papen H: **Soil-atmosphere exchange of N<sub>2</sub>O and NO in near-natural savanna and agricultural land in Burkina Faso (W. Africa).** *Ecosystems* 2008, **11**:582-600.
26. Meixner FX, Fickinger T, Marufu L, Serça D, Nathaus FJ, Makina E, Mukurumbira L, Andreae MO: **Preliminary results on nitric oxide emission from a southern African savanna ecosystem.** *Nutrient Cycling in Agroecosystems* 1997, **48**:123-138.
27. Olivier JGJ, Bouwman AF, Vandermaas CWM, Berdowski JJM: *Emission Database for Global Atmospheric Research (EDGAR).* Kluwer Academic Publishers; 1994:93-106.
28. EDGAR: Emission Database for Global Atmospheric Research; <http://edgar.jrc.ec.europa.eu/index.php> (accessed in April 2011).
29. Oenema O, Wrage N, Velthof GL, Van Groeningen JW, Dolfing J, Kuikman PJ: **Trends in global N<sub>2</sub>O emissions from animal production systems.** *Nutrient Cycling in Agroecosystem* 2005, **72**:51-65.
30. Rufino MC, Rowe EC, Delve RJ, Giller KE: **Nitrogen cycling efficiencies through resource-poor African crop-livestock system.** *Agriculture Ecosystems and Environment* 2006, **112**:261-282.
31. IPCC: *Intergovernmental Panel on Climate Change Guidelines for National Greenhouse Gas Inventories.* OECD; 1997.
32. Holst J, Liu C, Yao Z, Bruggemann N, Zheng X, Han X, Butterbach-Bahl K: **Importance of point sources on regional nitrous oxide fluxes in semi-arid steppe of Inner Mongolia, China.** *Plant and Soil* 2007, **296**:209-226.
33. FAO: *Gridded Livestock of the World.* Food and Agriculture Organization of the United Nations, FAO Animal Production and Health Division; 2007.
34. Velthof GL, Brader AB, Oenema O: **Seasonal variations in nitrous oxide losses from managed grasslands in The Netherlands.** *Plant and Soil* 1996, **181**:263-274.
35. Saggat S, Hedley CB, Giltrap DL, Lambie SM: **Measured and modelled estimates of nitrous oxide emission and methane consumption from a sheep-grazed pasture.** *Agriculture, Ecosystems & Environment* 2007, **122**:357-365.
36. Yamulki S, Jarvis S, Owen P: **Nitrous oxide emissions from excreta applied in a simulated grazing pattern.** *Soil Biology and Biochemistry* 1998, **30**:491-500.
37. Palm CA, Gachengo CN, Delve RJ, Cadish G, Giller KE: **Organic inputs for soil fertility management in tropical ecosystems: application of an organic resource database.** *Agriculture, Ecosystems & Environment* 2001, **83**:27-42.
38. Liu J, Liangzhi Y, Amini M, Obersteiner M, Herrero M, Zehnder AJB: **A high-resolution assessment on global nitrogen flows in cropland.** In *Proceedings of National Academy of Sciences of United States of America* 2010, **107**:8035-8040.
39. Titttonell P, Rufino MC, Janssen BH, Giller KE: **Carbon and nutrient losses during manure storage under traditional and improved practices in smallholder crop-livestock systems — evidence from Kenya.** *Plant and Soil* 2010, **328**:253-269.
40. Predotova M, Schlecht E, Buerkert A: **Nitrogen and carbon losses from dung storage in urban gardens of Niamey, Niger.** *Nutrient Cycling in Agroecosystems* 2010, **87**:103-114.
41. Palm CA, Alegre JC, Arevalo L, Mutuo PK, Mosier AR, Coe R: **Nitrous oxide and methane fluxes in six different land use systems in the Peruvian Amazon.** *Global Biogeochemical Cycles* 2002, **16**:1073.
42. Chikowo R, Mapfumo P, Nyamugafata P, Giller KE: **Mineral N dynamics, leaching and nitrous oxide losses under maize following two-year improved fallows on a sandy loam soil in Zimbabwe.** *Plant and Soil* 2004, **259**:315-330.
43. Galloway JN, Bekunda M, Cai Z, Erismann JW, Freney J, Howarth RW, Martinelli LA, Scholes MC, Seitzinger SP: **A preliminary assessment of changes in the global N cycle as result of anthropogenic influence.** *Third International Nitrogen Conference; Nanjing, China, October 12-16: 2004.* available at: <http://introgen.org>.
44. Beaulieu JJ, Tank JL, Hamilton SK, Wollheim WM, Hall RO, Mulholland PJ, Peterson BJ, Ashkenas LR, Cooper LW, Dahm CN *et al.*: **Nitrous oxide emission from denitrification in stream and river networks.** In *Proceedings of the National Academy of Sciences of United States of America* 2011, **108**:214-219 The paper addresses the nitrous oxide emissions from streams, rivers and estuaries as a major source of uncertainty of the N<sub>2</sub>O budget. Their estimate of stream and rivers nitrous oxide emissions is three times greater than estimated by the IPCC..
45. De Klein C, Novoa RSA, Ogle S, Smith KA, Rochette P, Wirth TC, McConkey BG, Mosier AR, Rypdal K, Walsh M, Williams SA: **N<sub>2</sub>O emissions from managed soils, and CO<sub>2</sub> emissions from lime and urea application.** In *2006 IPCC Guidelines for National Greenhouse Gas Inventories.* Edited by Eggleston S, Buendia L, Miwa K, Ngara T, Tanabe K. Japan: Institute for Global Environmental Strategies; 2006:1-54.



46. Kroeze C, Dumont E, Seitzinger SP: **Future trends in emissions of N<sub>2</sub>O from rivers and estuaries.** *Journal of Integrative Environmental Sciences* 2010, **7**:71-78.
47. Kroeze C, Dumont E, Seitzinger SP: **New estimates of global emissions of N<sub>2</sub>O from rivers, estuaries and continental shelves.** *Environmental Sciences* 2005, **2**:159-167.
48. Syakila A, Kroeze C, Slomp CP: **Neglecting sinks for N<sub>2</sub>O at the earth's surface: does it matter?** *Journal of Integrative Environmental Sciences* 2010, **7**:79-87.
49. Mosier AR, Kroeze C, Nevison C, Oenema O, Seitzinger S, Van Cleemput O: **Closing the global N<sub>2</sub>O budget: nitrous oxide emissions through the agricultural nitrogen cycle.** *Nutrient Cycling in Agroecosystems* 1998, **52**:225-248.
50. Alcamo J, Van Vuuren D, Cramer W, Alder J, Bennett E, Carpenter S, Christensen V, Foley J, Maerker M, Masui T *et al.*: **Changes in ecosystem services and their drivers across the scenarios.** In *Millennium Ecosystem Assessment. Volume 2: Scenarios.* Edited by Carpenter *et al.*: Island Press; 2005:297-373.
51. Carpenter SR, Pingali PL, Bennet EM: *Ecosystems and Human Well-being Scenarios, Findings of the Scenarios Working Group, Millennium Ecosystem Assessment.* London, Washington, Covello: Island Press; 2005.
52. Bouwman AF, Beusen AHW, Billen G: **Human alteration of the global nitrogen and phosphorus soil balances for the period 1970–2050.** *Global Biogeochemical Cycles* 2010:GB0A04 doi: 10.1029/2009GB003576.
53. Van Drecht G, Bouwman AF, Harrison J, Knoop JM: **Global nitrogen and phosphate in urban waste water for the period 1970–2050.** *Global Biogeochemical Cycles* 2009, **23**:GB0A03 doi: 10.1029/2009GB003458.
54. Bouwman AF, Kram T, Klein Goldewijk K (Eds): *Integrated Modelling of Global Environmental Change. An Overview of IMAGE 2.4.* Bilthoven: Netherlands Environmental Assessment Agency (MNP); 2006.
55. Crutzen PJ, Mosier AR, Smith KA, Winiwarter W: **N<sub>2</sub>O release from agro-biofuel production negates global warming reduction by replacing fossil fuels.** *Atmospheric Chemistry and Physics* 2008, **8**:389-395.
56. Davidson EA: **The contribution of manure and fertilizer nitrogen to atmospheric nitrous oxide since 1860.** *Nature Geoscience* 2009, **2**:659-662.
57. Hoben JP, Gehl RJ, Robertson GP, Millar N, Grace PR: **Nonlinear nitrous oxide (N<sub>2</sub>O) response to nitrogen fertilizer in on-farm corn crops of the US Midwest.** *Global Change Biology* 2011, **17**:1140-1152.
58. Van Groenigen JW, Velthof GL, Van Groenigen KJ, Van Kessel C: **Towards an agronomic assessment of N<sub>2</sub>O emissions: a case study for arable crops.** *European Journal of Soil Science* 2010, **61**:903-913.
59. Kimetu JM, Mugendi DN, Batiano A, Palm CA, Mutuo PK, Kihara J, Nandwa S, Giller KE: **Partial balance of nitrogen in a maize cropping system in humid nitisol of Central Kenya.** *Nutrient Cycling in Agroecosystems* 2006, **76**:261-270.
60. Millar N, Ndufa JK, Cadisch G, Baggs EM: **Nitrous oxide emissions following incorporation of improved-fallow residues in the humid tropics.** *Global Biogeochemical Cycles* 2004, **18**:GB1032 doi: 10.1029/2003GB002114.
61. Predotova M, Gebauer J, Diogo RVC, Schlecht E, Buerkert A: **Emissions of ammonia, nitrous oxide, and carbon dioxide from urban gardens of Niamey, Niger.** *Field Crops Research* 2010, **115**:1-8.
62. Dick J, Kaya B, Soutoura M, Skiba U, Smith R, Niang A, Tabo R: **The contribution of agricultural practices to nitrous oxide emissions in semi-arid Mali.** *Soil Use and Management* 2008, **24**:292-301.
63. FAO/IFA: *Global Estimates of Gaseous Emissions of NH<sub>3</sub>, NO and N<sub>2</sub>O from Agricultural Land.* Rome: Food and Agriculture Organization of the United Nations (FAO)/International Fertilizer Industry Association (IFA); 2001:106.
64. Levine JS, Winstead EL, Parsons DAB, Scholes M, Scholes RJ, Cofer WR, Cahoon DR, Sebacher DI: **Biogenic soil emissions of nitric oxide (NO) and nitrous oxide (N<sub>2</sub>O) from savannas in South Africa: the impact of wetting and burning.** *Journal of Geophysical Resources* 1996, **101**:23689-23697.
65. Chikowo R, Mapfumo P, Leffelaar PA, Giller KE: **Integrating legumes to improve N cycling on smallholder farms in Zimbabwe: resource quality, biophysical and environmental limitations.** *Nutrient Cycling in Agroecosystems* 2006, **76**:219-231.
66. Wulf S, Lehmann J, Zech W: **Emissions of nitrous oxide from runoff-irrigated and rainfed soils in semiarid north-west Kenya.** *Agriculture, Ecosystems & Environment* 1999, **72**:201-205.
67. Davidson EA, Vitousek PM, Matson PA, Riley R, Garciamendez G, Maass JM: **Soil emissions of nitric-oxide in a seasonally dry tropical forest of Mexico.** *Journal of Geophysical Resources* 1991, **96**:15439-15445.
68. Bouwman AF: **Direct emission of nitrous oxide from agricultural soils.** *Nutrient Cycling in Agroecosystems* 1996, **46**:53-70.
69. Del Grosso SJ, Ojima DS, Parton WJ, Stehfest E, Heistemann M, DeAngelo B, Rose S: **Global scale DAYCENT model analysis of greenhouse gas emissions and mitigation strategies for cropped soils.** *Global and Planetary Change* 2009, **67**:44-50.
70. Sanchez PA, Ahamed S, Carre F, Hartemink AE, Hempel J, Huising J, Lagacherie P, McBratney AB, McKenzie NJ, Mendonca-Santos MD *et al.*: **Digital soil map of the world.** *Science* 2009, **325**:680-681.
71. Sanchez PA, Ahamed S, Carre F, Hartemink AE, Hempel J, Huising J, Lagacherie P, McBratney AB, McKenzie NJ, de Lourdes Mendonca-Santos M *et al.*: **Digital soil map of the world.** *Science* 2009, **325**:680-681.